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Outdoor PM_{2.5}, Ambient Air Temperature, and Asthma Symptoms in the Past 14 Days among Adults with Active Asthma

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Running Title: PM_{2.5} and asthma exacerbations among adults

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Disclaimer

The findings and conclusions in this report are those of the authors and do not necessarily represent the official position of the Centers for Disease Control and Prevention.

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ABSTRACT

Background: Relationships between air quality and health are well-described, but little information is available about the joint associations between particulate air pollution, ambient temperature, and respiratory morbidity.

Objectives: To evaluate associations between concentrations of particulate matter \leq 2.5 microns in diameter (PM_{2.5}) and exacerbation of existing asthma and modification of the associations by ambient air temperature.

Methods: Data from 50,356 adult 2006–2010 Asthma Call-back Survey respondents were linked by interview date and county of residence to estimates of daily averages of PM_{2.5} and maximum air temperature. Associations between 14-day average PM_{2.5} and the presence of any asthma symptoms during the 14 days leading up to and including the interview date were evaluated using binomial regression. We explored variation by air temperature using similar models, stratified into quintiles of the 14-day average maximum temperature.

Results: Among adults with active asthma, 57.1% reported asthma symptoms within the past 14 days and 14-day average $PM_{2.5} \ge 7.07 \,\mu g \cdot m^{-3}$ was associated with an estimated 4 to 5% higher asthma symptom prevalence. In the range of 4.00 to 7.06 $\mu g \cdot m^{-3}$ of $PM_{2.5}$, each $\mu g \cdot m^{-3}$ increase was associated with a 3.4% (95% confidence interval: 1.1, 5.7) increase in symptom prevalence; across categories of temperature from 1.1 to 80.5°F, each $\mu g \cdot m^{-3}$ increase was associated with increased symptom prevalence (1.1–44.4°F: 7.9%; 44.5–58.6°F: 6.9%; 58.7–70.1°F: 2.9%; 70.2–80.5°F: 7.3%).

Conclusions: These results suggest that each unit increase in PM_{2.5} may be associated with an increase in the prevalence of asthma symptoms, even at levels as low as 4.00 to 7.06 μ g·m⁻³.

INTRODUCTION

Ambient particulate matter (PM) pollution accounted for an estimated 3.1 million deaths and 3% of global disability-adjusted life years in 2010 (Lim et al. 2012). Health studies examining the effects of PM exposures have identified links between PM and new-onset asthma (Young et al. 2014), respiratory symptoms (Balmes et al. 2014; Young et al. 2014), hospitalizations and emergency department visits (Bell et al. 2008; Belleudi et al. 2010; Dominici et al. 2006), and death (Dominici et al. 2000; Samet et al. 2000; Zanobetti and Schwartz 2009) among adults. The biologic effect of particulate air pollution on respiratory health is determined largely by the size and composition of the particulate air pollution, deposition of particles in the respiratory tract, and the immunologic response to the particles (Koren 1995). Mechanisms by which exposure to particulate air pollution may exacerbate respiratory health among individuals with asthma include oxidative stress, airway inflammation, and hyperresponsiveness of the airways (Bernstein et al. 2004; Li et al. 2003; Nel et al. 2001). In recognition of the public health importance of the effects of particulate air pollution exposures, standards such as the California Air Resources Board Air Quality Standards, European Union Directives, U.S. National Ambient Air Quality Standards, and World Health Organization Air Quality Guidelines have been used around the world to establish ambient air quality standards (Vahlsing and Smith 2012).

A growing body of epidemiologic literature also provides initial evidence that exacerbations of adult asthma may be associated with ambient meteorological conditions, particularly with temperature extremes (Beard et al. 2012; Fitzgerald et al. 2014; Michelozzi et al. 2009). However, the mechanisms through which outdoor temperature exposures may plausibly affect the respiratory tract remain unclear. Proposed mechanisms include an effect of temperature or

other ambient conditions on dehydration and hyperosmolarity of the airways, which may directly induce exacerbations of asthma by triggering bronchoconstriction (Anderson and Daviskas 2000). Proposed indirect effects focus on exposures associated with temperature; for example, well-described associations between ozone (O₃) and airway inflammation, particularly among adults with asthma (Hernandez et al. 2010; Khatri et al. 2009; Koren 1995), raise the possibility that the observed associations between temperature and exacerbations of asthma may be attributed, at least in part, to changes in the production of ground-level O₃ that are correlated with temperature and other meteorological factors (Baur et al. 2004; Camalier et al. 2007). Higher temperatures are also associated with increased air pollutant emissions; for example, hot spells can lead to escalated use of air conditioning thereby increasing electrical demand on electricity-generating units, which can in turn lead to increased oxides of nitrogen (NO_x) emissions on hot days (He et al. 2013). The proposed pathways between temperature and exacerbations of asthma may also be modified by adaptive behaviors such as air conditioning use or avoidance of outdoor activities (Andrade et al. 2011). In the absence of clearly delineated mechanisms by which temperature is associated with mortality across the entire range of ambient outdoor temperatures, numerous recent studies have been designed to evaluate interaction between particulate air pollution and temperature to affect mortality, including respiratory, cardiovascular, and all-cause mortality (Breitner et al. 2014; Li et al. 2011; Nawrot et al. 2007; Park et al. 2011).

To date, however, little information is available about the joint associations between particulate air pollution, ambient temperature, and respiratory morbidity. Data from the Centers for Disease Control and Prevention's Environmental Public Health Tracking Network and the Behavioral

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Risk Factor Surveillance System (BRFSS) adult Asthma Call-back Survey provide a unique opportunity to explore further such a possibility by evaluating associations between ambient PM and asthma symptoms among adults with asthma in a large and geographically diverse sample of adults. In this study, we combined modeled county-level estimates of ambient concentrations of PM ≤ 2.5 microns in aerodynamic diameter (PM_{2.5}), ambient concentrations of O₃, precipitation, and air temperature with individual-level characteristics of adults with asthma to evaluate associations between PM_{2.5} and asthma exacerbations in the United States during the period 2006–2010 and to describe the extent to which the observed associations may vary by air temperature. To accomplish these objectives, our study included two major sets of analyses. First, we conducted a main analysis of associations between PM_{2.5} and asthma exacerbations in which PM_{2.5} was evaluated using quartiles, linear splines, and a single continuous measure. Second, we stratified the main analysis into quintiles of air temperature to explore effect modification of the main results by ambient air temperature.

METHODS

Asthma Call-back Survey

We conducted these analyses using data from the 2006–2010 BRFSS adult Asthma Call-back Surveys. The BRFSS is a state-level survey of the adult civilian, noninstitutionalized population aged 18 years and older conducted annually in the United States (Centers for Disease Control and Prevention 2009). BFRSS survey interviews are conducted throughout year (i.e., January through December). The Asthma Call-back Survey is a follow-up telephone survey conducted approximately two weeks after the BRFSS among respondents who indicated that they have ever had asthma. Respondents reported ever having had asthma by responding "yes" to the following

question: "have you ever been told by a doctor, nurse, or other health professional that you had asthma?" The Asthma Call-back Survey was administered to 10,801 respondents in 2006; 15,245 respondents in 2007; 15,007 in 2008; 15,403 in 2009; and 17,753 in 2010. In the participating areas included in our analysis, the Council of American Survey and Research Organization response rates for the Asthma Call-back Surveys ranged from 41 to 71% in 2006, 36 to 72% in 2007, 35 to 68% in 2008, 36 to 66% in 2009, and 31 to 67% in 2010 (Mirabelli et al. 2014; National Asthma Control Program 2011a, b, 2012). The Asthma Call-back Survey is exempt from Institutional Review Board (IRB) review at the Centers for Disease Control and Prevention (CDC); state-specific IRB requirements apply to each of the participating states, the District of Columbia, and Puerto Rico. The protocol for the present analysis was reviewed and determined to be exempt from IRB review at CDC.

Study Sample

For this analysis, we present results based on a sample of 50,356 Asthma Call-back Survey respondents. The sample was generated by pooling data collected from Asthma Call back-Surveys conducted in 2006, 2007, 2008, 2009, and 2010. The pooled sample included 74,209 respondents from 42 geographic areas of the United States (40 states, the District of Columbia, and Puerto Rico). We limited our analysis to respondents with active asthma (n=56,509; 76%). We then excluded respondents for whom county- and date-linked air quality data were unavailable (n=3,200, including all respondents from Alaska, Hawaii, and Puerto Rico) and with missing data for the asthma symptom questions and covariates included in our final analysis (n=1,998). Finally, to reduce the influence of the few observations at the lower and upper tails of the distribution of 14-day average PM_{2.5}, we excluded from our analyses 418 (1%) observations

with PM_{2.5}<4.00 micrograms per cubic meter ($\mu g \cdot m^{-3}$) and 537 (1%) observations PM_{2.5}>20.00 $\mu g \cdot m^{-3}$. The values of 4.00 $\mu g \cdot m^{-3}$ and 20.00 $\mu g \cdot m^{-3}$ were selected as minimum and maximum values using the 1st and 99th percentile values of 4.09 $\mu g \cdot m^{-3}$ and 20.12 $\mu g \cdot m^{-3}$, respectively, of the distribution of 14-day average PM_{2.5}, rounded to the nearest integer.

County-level Estimates of Environmental Variables

For this analysis, we linked data from the Asthma Call-back Survey with county-level estimates of environmental variables. In the United States, counties (or equivalent entities such as boroughs, parishes, and independent cities) are the legally defined political and administrative units within each state (U.S. Department of Commerce Bureau of the Census 1994). Among the geographic areas included in our analysis, the number of counties (or equivalent entities, hereafter referred to as 'counties') ranged from 1, in the District of Columbia, to 254, in Texas; in total, we linked environmental data from 2,253 U.S. counties for this analysis. Daily estimates of PM_{2.5} and O₃ generated using a Bayesian space-time Downscaler fusion model (Berrocal et al. 2012). The Downscaler modeling approach combines output from the Community Multi-scale Air Quality (CMAQ) model (Byun and Schere 2006; Foley et al. 2010) with measurements from the U.S. Environmental Protection Agency's Air Quality System (Centers for Disease Control and Prevention 2013; U.S. Environmental Protection Agency) to yield air quality predictions at specific geographic locations. The CMAQ model is a multi-pollutant, multi-scale chemical transport model used to generate air quality predictions at user-defined spatio-temporal scales, taking into account land use, chemical transport, chemistry, emission processes, land use, and weather (Byun and Schere 2006; Foley et al. 2010). Descriptions of the theory, development, and initial evaluation of the Downscaler model have been published previously (Berrocal et al.

2010a, b, 2012). Daily predictions of 24-hour average PM_{2.5} concentrations in μg·m⁻³ and daily maximum 8-hour average O₃ concentrations in parts per billion (ppb) were generated at 2010 U.S. census tract centroid locations using the Downscaler software (Heaton et al. 2012). In addition, daily county-level estimates of PM_{2.5} and O₃ were generated using a population-weighted approach in which census tract population counts were used to weight daily census tract-level PM_{2.5} and O₃ predictions (Ivy et al. 2008; Vaidyanathan et al. 2013). County-level estimates of precipitation in millimeters (mm) and ambient air temperature in degrees Fahrenheit (°F) were generated using meteorological predictions from the North American Land Data Assimilation System Phase 2 model (Mitchell et al. 2004) at 0.125 degree spatial resolution (i.e., approximately 14x14 kilometers).

At the time these analyses were conducted, estimates of environmental variables were not available for interview dates in 2011 and, as noted above, all survey respondents for whom environmental data were missing were excluded from analysis. For each of the remaining respondents in our analysis, we assigned 14-day estimates of PM_{2.5}, O₃, precipitation, and temperature using his/her county of residence and the 14-day period leading up to and including the date on which the Asthma Call-back Survey interview was conducted.

For PM_{2.5}, we assigned a county-level average of the daily 24-hour average PM_{2.5} concentrations estimated for the 14-day period leading up to and including the day of the interview. Without an a priori hypothesis about the best metric of PM_{2.5} to use, we categorized the distribution of 14-day county-level average PM_{2.5} using three metrics: (1) quartiles (quartile 1: $4.00-7.06 \,\mu g \cdot m^{-3}$; quartile 2: $7.07-8.97 \,\mu g \cdot m^{-3}$; quartile 3: $8.98-11.36 \,\mu g \cdot m^{-3}$; and quartile 4: $11.37-19.98 \,\mu g \cdot m^{-3}$),

(2) linear spline segments specified with knots between quartiles, and (3) a single continuous measure of $PM_{2.5}$.

Estimates of O_3 , precipitation, and temperature were each assigned using county-level averages of values estimated for the 14-day period leading up to and including the day of the interview. For O_3 , we assigned a county-level average of the daily 8-hour maximum O_3 concentrations estimated for the 14-day period leading up to and including the day of the interview. Our final models include O_3 parameterized using deciles of the distribution, with the 6^{th} decile (range: 38.6-41.4 ppb) as the referent category; the 6^{th} decile was selected as the referent because it included the mean value of the distribution of O_3 (mean: 38.7 (standard deviation (SD): 9.7); range: 10.3-83.8 ppb). Our final models include precipitation in categories: 0.0 mm, 0.1-1.1 mm, 1.2-3.0 mm, and ≥ 3.1 mm, with 0.0 mm as the referent category; estimated precipitation greater than 0.0 mm was categorized as tertiles of the distribution. As with O_3 , our final models include temperature using deciles of the distribution of maximum temperature, with the 5^{th} decile as the referent (range: 58.8-64.6°F [14.9-18.1°C]), also selected as the referent category because it included the mean value of the distribution of temperature (mean (SD): 62.9°F (19.3) [17.2°C (10.7)]; range: 1.1-112.4°F [-17.2-44.7°C]).

Asthma

As in previous analyses (Mirabelli et al. 2014), respondents were categorized as having active asthma if they reported that at least one of the following occurred during the past 12 months: talked to a doctor or other health professional about [his/her] asthma, took asthma medication, or experienced any symptoms of asthma. As an indicator of the presence of asthma symptoms in the

past two weeks, we used responses to the following questionnaire item: "during the past two weeks, on how many days were you completely symptom-free – that is, no coughing, wheezing, or other symptoms of asthma?" Respondents who reported being symptom-free on fewer than 14 of the past 14 days were categorized as having asthma symptoms in the past two weeks.

Other Covariates

Demographic covariates used in this analysis include age, educational attainment, race/ethnicity, and sex. Cigarette smoking status was categorized as current smoker, former smoker, and lifetime nonsmoker. States represented in our final sample were grouped into eight U.S. climate regions classified by the National Climatic Data Center (Karl and Koss 1984; National Climatic Data Center. National Oceanic and Atmospheric Administration): central (Illinois, Indiana, Missouri, Ohio, West Virginia), east north central (Iowa, Michigan, Wisconsin), northeast (Connecticut, District of Columbia, Maine, Maryland, Massachusetts, New Hampshire, New Jersey, New York, Pennsylvania, Rhode Island, Vermont), northwest (Oregon, Washington), south (Kansas, Louisiana, Mississippi, Oklahoma, Texas), southeast (Alabama, Florida, Georgia, Virginia), southwest (Arizona, Colorado, New Mexico, Utah), west (California, Nevada), and west north central (Montana, Nebraska, North Dakota). The urbanicity of each county in our final sample was categorized as rural, suburban, or urban using 2013 Economic Research Service rural-urban continuum codes published by the U.S. Department of Agriculture (U.S. Department of Agriculture 2013), with counties not adjacent to a metro area categorized as rural, counties adjacent to a metro area categorized as suburban, and counties in metro areas categorized as urban.

Statistical Analysis

Demographic characteristics and smoking status are presented for the sample of respondents and the weighted population estimate. Weighted population estimates were generated using adjusted sampling weights applied to account for BRFSS and Asthma Call-back Survey nonresponse and unequal sampling probabilities. Annual Asthma Call-back Survey sampling weights were provided with the Asthma Call-back Survey data. Because we pooled data collected from 2006 through 2010 and because the number of geographic areas with Asthma Call-back Survey respondents varied from year to year, we re-adjusted the sampling weights in each geographic area by dividing the annual Asthma Call-back Survey weights by the number of years for which data were available (Mirabelli et al. 2014). All descriptive analyses were performed using survey procedures (e.g., PROC SURVEYFREQ) for the analysis of complex survey data in SAS (SAS Institute, Inc.).

Using additive binomial models specified with an identity link, we estimated the prevalence of asthma symptoms during the past 14 days for the entire weighted population estimate and across categories of individual- and county-level characteristics, accounting for complex survey sampling. Associations between PM_{2.5} and asthma symptoms during the past 14 days were evaluated using PROC SURVEYREG in SAS (SAS Institute, Inc.), with standard errors generated by adapting a two-step approach developed and published by Natarajan and colleagues designed to generate robust variance estimates to fit the binomial error distribution of our data (Natarajan et al. 2008; Slade et al. 2012). Because we conduced our analyses using statistical software designed to analyze complex survey data, all models accounted for the state-based sampling approach and our pooling of five years of survey data. Models were adjusted for

individual-level covariates (age, educational attainment, race/ethnicity, sex, and smoking status) and county-level covariates (O₃, precipitation, region, temperature, and urbanicity). Following these analyses, we conducted two additional analyses in which we replaced the measure of ambient air temperature, initially based on the mean of distribution of daily maximum temperatures during the 14-day period, with the median of the distribution and with the mean apparent temperature during the 14-day period, respectively.

To evaluate the extent to which the observed associations between $PM_{2.5}$ and asthma symptoms in the past 14 days vary by temperature, we stratified adjusted models into quintiles of the distribution of daily maximum ambient air temperature (1.1–44.4°F [-17.2–6.8°C], 44.5–58.6°F [6.9–14.8°C], 58.7–70.1°F [14.9–21.1°C], 70.2–80.5°F [21.2–26.9°C], and 80.6–112.4°F [27.0–44.7°C]).

Measures of association are presented as adjusted percent differences (PDs) with 95% confidence intervals (CIs). For analyses in which PM_{2.5} was evaluated using quartiles, PDs are interpreted as arithmetic differences in the prevalence of asthma symptoms in PM_{2.5} quartiles 2, 3, and 4 compared to the prevalence of asthma symptoms in referent quartile 1. For analyses of linear spline segments and analyses in which PM_{2.5} was include as a single continuous measure, PDs are interpreted as the change in prevalence of asthma symptoms per one $\mu g \cdot m^{-3}$ increase in PM_{2.5}. All analyses were conducted using SAS version 9.3 (SAS Institute, Inc., Cary, North Carolina).

RESULTS

Characteristics of the 50,356 adults with active asthma and the weighted population estimate are shown in Table 1. Among adults with active asthma, an estimated 57.1% reported asthma symptoms within the past 14 days. Variations in percentages of adults reporting asthma symptoms within the past 14 days were observed across categories of respondent characteristics as well as by U.S. climate region and county-level urbanicity.

Summary statistics describing the distributions of $PM_{2.5}$, O_3 , temperature, and precipitation as well as Pearson's correlation coefficients indicating correlations between the four measures are shown in Table 2. Despite the small magnitudes of the correlations, all pairwise correlation coefficients were statistically significant at α <0.05. The distributions of $PM_{2.5}$, O_3 , precipitation, and temperature for the weighted population estimate of nearly 18.0 million adults with active asthma and the estimated percentages of adults with active asthma reporting asthma symptoms in the past 14 days across categories of each environmental measure are shown in Figure 1. Across quartiles of $PM_{2.5}$, the percentages of adults reporting asthma symptoms in the past 14 days were $54.2\% \pm 1.2$ in quartile 1, $58.4\% \pm 1.0$ in quartile 2, $58.3\% \pm SE$: 1.0 in quartile 3, and $56.7\% \pm 1.1$ in quartile 4.

Table 3 shows estimates of the difference in the percentage of adults with active asthma who experienced symptoms in the past 14 days, generated using unadjusted models and models adjusted for individual- and county-level covariates. Broadly speaking, estimates generated using models adjusted for individual-level covariates were similar to those generated using unadjusted models and estimates generated using models adjusted for county-level covariates were similar

to those generated using fully adjusted models. In fully adjusted models, when $PM_{2.5}$ exposure was parameterized using indicators of $PM_{2.5}$ quartile, the estimated prevalences of asthma symptoms in the past 14 days were typically 4 to 5% higher in quartiles 2, 3, and 4, relative to the prevalence in quartile 1. In models adjusted for individual- and county-level covariates in which $PM_{2.5}$ was parameterized as four linear spline segments, percent differences in 14-day symptom prevalence per 1 μ g·m⁻³ increase in $PM_{2.5}$ were 3.4% (95% CI: 1.1, 5.7) in segment 1, 0.9 (95% CI: -1.3, 3.1) in segment 2, -0.1 (95% CI: -1.6, 1.5) in segment 3, and 0.3 (95% CI: -0.6, 1.1) in segment 4. These PDs indicate an estimated increase of 3.4% of adults with active asthma reporting symptoms in the past 14 days with each 1 μ g·m⁻³ unit increase in $PM_{2.5}$ between 4.00 and 7.06 μ g·m⁻³ and minimal change per unit increase from 7.07 to 19.98 μ g·m⁻³. In contrast, when we considered adjusted associations with each 1 μ g·m⁻³ unit increase in $PM_{2.5}$ across the entire distribution of $PM_{2.5}$, our models generated a per unit increase in symptom prevalence of 0.5% (95% CI: 0.1, 0.9).

These results were robust to changes in the measure of ambient air temperature. Our main analyses included ambient air temperature parameterized using the mean of the distribution of 14-day average daily maximum temperatures. Replacing this measure with variables parameterized using the median of the distribution or the mean apparent temperature generated estimates of effect that were nearly identical in magnitude and precision (data not shown).

Variations in the estimates across categories of air temperature are shown in Figure 2. When PM_{2.5} exposure was modeled using quartiles, estimated differences in the prevalence of asthma symptoms in the past 14 days were similar across all categories of temperature and highest in the

70.2–80.5°F range (quartile 2: 9.4%; 95% CI: 3.1, 15.7; quartile 3: 11.0%; 95% CI: 4.4, 17.6; and quartile 4: 6.1%; 95% CI:-1.1, 13.3) (Table S1). When PM_{2.5} exposure was parameterized as linear spline segments, the pattern observed in our main analysis was also observed across quintiles of temperature. Spline segment 1 was associated with positive point estimates of the PD per μg·m⁻³ of PM_{2.5} across categories of temperature ranging from 1.1 to 80.5°F (1.1–44.4°F: 7.9%; 44.5–58.6°F: 6.9; 58.7–70.1°F: 2.9%; 70.2–80.5°F: 7.3%). Per μg·m⁻³ changes in symptom prevalence in spline segments 2 through 4 were consistent with the null value of 0.0%. Minimal variation was observed across categories of temperature when PM_{2.5} was evaluated as a single continuous measure.

DISCUSSION

Our findings suggest that among adults with asthma the prevalence of self-reported asthma symptoms during the past 14 days was 4 to 5% higher among respondents with a 14-day average concentration of $PM_{2.5}$ greater than 7.07 $\mu g \cdot m^{-3}$, compared to the prevalence among respondents with 14-day averages in the range of 4.00 to 7.06 $\mu g \cdot m^{-3}$. In this lowest category of $PM_{2.5}$, each 1 $\mu g \cdot m^{-3}$ increase in $PM_{2.5}$ was associated with a 3% increase in the prevalence of asthma symptoms. In combination, our results raise the possibility that among individuals with asthma, exacerbations may begin to increase at relatively low levels of ambient $PM_{2.5}$. Stratification of these results across categories of temperature suggests that these findings may be driven largely by effects observed at temperatures below $80.6^{\circ}F$.

These results expand on previous research that evaluated relationships between $PM_{2.5}$ and asthma symptoms across quartiles of the distribution of $PM_{2.5}$ (Mirabelli et al. 2015), by exploring

additional metrics of PM_{2.5} and refining the statistical models used to estimate the exposureoutcome relationships. Previously, we reported on quartiles of the PM_{2.5} distribution and found higher percentages of adults with asthma symptoms in the past 14 days in quartiles 2, 3, and 4 compared to quartile 1 (Mirabelli et al. 2015). In considering relationships between PM_{2.5} and asthma symptoms using four linear spline segments, the present analysis supports and expands our earlier findings by suggesting that each unit increase in PM_{2.5} may be associated with a measurable increase in the prevalence of asthma symptoms in the past 14 days, even at levels as low as 4.00 to 7.06 ug·m⁻³. Indeed, using results from our fully-adjusted linear spline segment model, PD estimates generated at the median of each spline segments support the results of our analysis using quartiles. For example, where models using quartiles of PM_{2.5} generated PDs of 4.4%, 4.7%, and 4.9% in quartiles 2, 3, and 4, respectively, models using linear spline segments generated PDs of 4.5%, 5.2%, and 5.7% (data not shown). Both models suggest that among individuals with asthma, exacerbations may begin to increase at levels of PM_{2.5} below 7.07 μg·m⁻³. The strength of this association at levels between 4.00 and 7.06 μg·m⁻³ would have been missed had we only evaluated the relationship between PM_{2.5} and asthma symptoms using a single continuous measure of PM_{2.5}, which generated a 0.5% increase in the prevalence of asthma symptoms with each 1 ug·m⁻³ unit increase in PM_{2.5}.

Currently, there are few population-based studies of $PM_{2.5}$ and asthma morbidity outcomes with which to contrast our results. Following an evaluation of associations between daily minimum temperature and respiratory hospitalizations and emergency department visits, investigators reported that ambient concentrations of $PM \le 10$ microns in aerodynamic diameter (PM_{10}) modified the association between temperature and respiratory hospitalizations, but not

emergency department visits, and that the observed associations with increasing temperatures were more pronounced when concentrations of PM_{10} also increased (Ren et al. 2006). In combination with convincing evidence of the effects of ambient PM exposures on the respiratory health of adults (HEI Panel on the Health Effects of Traffic-Related Air Pollution 2010), our findings that the magnitudes of the associations between $PM_{2.5}$ and asthma symptoms are not constant across the range of $4.00-19.98~\mu g\cdot m^{-3}$ of $PM_{2.5}$ and that the associations vary across categories of ambient air temperature support and extend previous evidence that ambient particulate matter pollution and temperature may act jointly to affect respiratory health in manners that may be evident before death (Ren et al. 2006)

Several aspects of our study merit careful consideration when interpreting our findings. First, a comparison of results shown in Table 1 suggests that the relationship between PM_{2.5} and the prevalence of asthma symptoms in the past 14 days may be confounded by county-level factors, including O₃, precipitation, and temperature. Despite including these and other county-level covariates in our analysis, our results may be affected by residual confounding, including within categories of ambient air temperature or by geographic differences within climate regions. We were unable to stratify our results into smaller categories of air temperature. Earlier research designed to estimate the effects of ambient air pollutant exposures on health initially suggested that statistical models that include one or two weather variables, such as temperature and humidity, may not be adequate to control fully for the effects of weather (Pope and Kalkstein 1996). From subsequent research designed to evaluate the extent to which statistical models of the relationship between particulate pollution and mortality that simultaneously account for the effects of weather or temperature may yield biased results if incorrect metrics of weather or

temperature are used, investigators reported finding little evidence that the choice of metrics influenced findings (Samet et al. 1998). Results generated from our main models, which included temperature and O₃ using indicators for deciles of the distributions and precipitation using indicators for tertiles of the distribution, may not have adequately accounted for the complex relationship between air quality, weather, and health, but were not notably changed when we considered alternative parameterizations of O₃, temperature, and precipitation (not shown). Improvements in our understanding of the complex relationships between air quality, weather, and health would improve our ability to incorporate these relationships, including nonlinear relationships and relationships at temperature extremes (Pope and Kalkstein 1996), into our analyses of the relationship between air quality and asthma exacerbations.

Second, modeled exposures provide estimates of PM_{2.5}, O₃, precipitation, and temperature for geographic areas in which respondents live. Assigning county-level exposure measures may have resulted in exposure misclassification if respondents true exposures were markedly different from those assigned to the counties in which they lived. The modeling approach applied in our study advantageously provided estimates for geographic areas without adequate measurements of air pollutants or meteorological parameters. However, uncertainties associated with estimates of exposures derived using modeled data could result in exposure misclassification, which is not incorporated into our statistical models, and we were unable to take into account a wide range of other air pollutants or PM of other size fractions (e.g., course or ultrafine). When concentrations of multiple air pollutants are highly correlated, results generated from models that do not include each of the multiple pollutants may be confounded by the effects of other pollutants not considered and we are unable to evaluate the extent to which our findings may represent

relationships between other air pollutants and exacerbations of asthma. Nonetheless, the integration of data from a large, population-based survey of adults with asthma with environmental data estimated at the county-level enabled us to evaluate the relationship between county-level PM_{2.5} and asthma symptoms in a large population of U.S. adults.

Our analyses do not account for indoor exposures, filtration of indoor air or air exchange rates, occupational exposures, pollen concentrations, or other factors affecting personal exposure. We do not have information about the extent to which respondents may have been included in two or more annual sample populations. Our analyses do not take into account the timing, including day of the week, month, or season, of the asthma call-back survey interview. In these data, correlations between interview month, U.S. climate region, and temperature prevented our statistical models from converging when all three were included. Similarly, imprecise estimates within each state prevented us from stratifying our analysis by state. Using these data, we were unable to evaluate temporal relationships between exposures and asthma symptoms within the 14 day period, therefore we cannot draw conclusions about the extent to which symptoms may have preceded or followed peak exposures. Our analyses also do not incorporate information about the changes in exposure to outdoor air or changes in behaviors that may occur at temperature extremes. If any changes, such as closing windows, using air conditioning, and staying indoors during hot weather, reduce exposures to ambient air pollutants, then our observation of attenuated associations at temperatures >80.6°F is unsurprising. In addition to the direct effects of exposure to ambient particulate matter pollution, susceptibility to the effects of exposure may be linked to other factors associated with asthma, such as diet and obesity, indoor air quality, medication use, socioeconomic status, and stress (Guarnieri and Balmes 2014). Some

populations, such as those living in neighborhoods located near highways or affected by industrial air emissions, may experience exposures higher than county averages and be more vulnerable to the effects of poor air quality. Additional information about other factors (e.g., pollen and other allergens) associated with exacerbation of asthma and about the frequency and severity of the asthma exacerbations would improve our ability to draw conclusions about the role of exposures to PM_{2.5} in the ambient air.

Despite these limitations, our findings extend our current understanding of the health effects of PM_{2.5} by considering self-reported asthma exacerbations, rather than more severe outcomes such as hospital encounters or mortality. Health effects associated with PM_{2.5} exposures are wideranging and may have measureable impacts on outcomes not considered in this analysis, including indicators of asthma severity, symptom frequency, medication use, functional consequences of asthma, as well as other cardiovascular and respiratory conditions. Our results suggest that the relationship between PM_{2.5} and asthma may not be constant across the entire range of PM_{2.5} concentration and, as a consequence, when the exposure of interest is PM_{2.5} in the ambient air, changes in population-level metrics of asthma exacerbations may occur most noticeably in the range of 4.00 to 7.06 μg·m⁻³. These findings provide novel information about the importance of county-level air quality for the nearly 18 million adults with asthma represented by the sampled survey population of Asthma Call-back Survey respondents.

REFERENCES

Anderson SD, Daviskas E. 2000. The mechanism of exercise-induced asthma is... J Allergy Clin Immunol 106:453-459.

Andrade H, Alcoforado MJ, Oliveira S. 2011. Perception of temperature and wind by users of public outdoor spaces: relationships with weather parameters and personal characteristics. Int J Biometeorol 55:665-680.

Balmes JR, Cisternas M, Quinlan PJ, Trupin L, Lurmann FW, Katz PP, et al. 2014. Annual average ambient particulate matter exposure estimates, measured home particulate matter, and hair nicotine are associated with respiratory outcomes in adults with asthma. Environ Res 129:1-10.

Baur D, Saisana M, Schulze N. 2004. Modelling the effects of meteorological variables on ozone concentration-a quantile regression approach. Atmospheric Environment 38:4689-4699.

Beard JD, Beck C, Graham R, Packham SC, Traphagan M, Giles RT, et al. 2012. Winter temperature inversions and emergency department visits for asthma in Salt Lake County, Utah, 2003-2008. Environ Health Perspect 120:1385-1390.

Bell ML, Ebisu K, Peng RD, Walker J, Samet JM, Zeger SL, et al. 2008. Seasonal and regional short-term effects of fine particles on hospital admissions in 202 US counties, 1999-2005. Am J Epidemiol 168:1301-1310.

Belleudi V, Faustini A, Stafoggia M, Cattani G, Marconi A, Perucci CA, et al. 2010. Impact of fine and ultrafine particles on emergency hospital admissions for cardiac and respiratory diseases. Epidemiology 21:414-423.

Bernstein JA, Alexis N, Barnes C, Bernstein IL, Nel A, Peden D, et al. 2004. Health effects of air pollution. J Allergy Clin Immunol 114:1116-1123.

Berrocal VJ, Gelfand AE, Holland DM. 2010a. A Spatio-Temporal Downscaler for Output From Numerical Models. J Agric Biol Environ Stat 15:176-197.

Berrocal VJ, Gelfand AE, Holland DM. 2010b. A bivariate space-time downscaler under space and time misalignment. Ann Appl Stat 4:1942-1975.

Berrocal VJ, Gelfand AE, Holland DM. 2012. Space-time data fusion under error in computer model output: an application to modeling air quality. Biometrics 68:837-848.

Breitner S, Wolf K, Devlin RB, Diaz-Sanchez D, Peters A, Schneider A. 2014. Short-term effects of air temperature on mortality and effect modification by air pollution in three cities of Bavaria, Germany: a time-series analysis. Sci Total Environ 485-486:49-61.

Byun D, Schere KL. 2006. Review of the governing equations, computational algorithms, and other components of the Models-3 Community Multiscale Air Quality (CMAQ) modeling system. Applied Mechanics Reviews 59:51-77.

Camalier L, Cox W, Dolwick P. 2007. The effects of meteorology on ozone in urban areas and their use in assessing ozone trends. Atmospheric Environment 41:4127-4137.

Centers for Disease Control and Prevention. 2009. Behavioral Risk Factor Surveillance System. Available: http://www.cdc.gov/asthma/survey/brfss.html [accessed January 23 2013].

Centers for Disease Control and Prevention. 2013. Outdoor air: monitor + modeled air data. Available: http://ephtracking.cdc.gov/showAirMonModData.action [accessed September 11 2014].

Dominici F, Samet JM, Zeger SL. 2000. Combining evidence on air pollution and daily mortality from the 20 largest US cities: a hierarchical modeling strategy. J R Statist Soc A 163:263-302.

Dominici F, Peng RD, Bell ML, Pham L, McDermott A, Zeger SL, et al. 2006. Fine particulate air pollution and hospital admission for cardiovascular and respiratory diseases. JAMA 295:1127-1134.

Fitzgerald EF, Pantea C, Lin S. 2014. Cold spells and the risk of hospitalization for asthma: New York, USA 1991-2006. Lung 192:947-954.

Foley KM, Roselle SJ, Appel Kw, Bhave PV, Pleim JE, Otte TL, et al. 2010. Incremental testing of the Community Multiscale Air Quality (CMAQ) modeling system version 4.7. Geosci Model Dev 3:205-226.

Guarnieri M, Balmes JR. 2014. Outdoor air pollution and asthma. Lancet 383:1581-1592.

He H, Hembeck L, Hosley K, Canty T, Salawitch RJ, Dickerson R. 2013. High ozone concentrations on hot days: the role of electric power demand and NO_x emissions. Geophysical Research Letters 40:5291-5294.

Heaton M, Holland DM, Leininger T. 2012. User's manual for Downscaler fusion software. (TIP#12-017). EPA/600/C-12/002. U.S. Environmental Protection Agency, Washington, D.C.

HEI Panel on the Health Effects of Traffic-Related Air Pollution. 2010. Traffic-related air pollution: a critical review of the literature on emissions, exposure, and health effects. HEI Special Report 17. Health Effects Institute, Boston, MA.

Hernandez ML, Lay JC, Harris B, Esther CR, Jr., Brickey WJ, Bromberg PA, et al. 2010. Atopic asthmatic subjects but not atopic subjects without asthma have enhanced inflammatory response to ozone. J Allergy Clin Immunol 126:537-544 e531.

Ivy D, Mulholland JA, Russell AG. 2008. Development of ambient air quality population-weighted metrics for use in time-series health studies. J Air Waste Manag Assoc 58:711-720.

Karl TR, Koss WJ. 1984. Regional and national monthly, seasonal, and annual temperature weighted by area, 1895-1983. Historical Climatology Series 4-3, 1984. National Climatic Data Center, Asheville, NC. pp. 38.

Khatri SB, Holguin FC, Ryan PB, Mannino D, Erzurum SC, Teague WG. 2009. Association of ambient ozone exposure with airway inflammation and allergy in adults with asthma. J Asthma 46:777-785.

Koren HS. 1995. Associations between criteria air pollutants and asthma. Environ Health Perspect 103 (Suppl 6):235-242.

Li G, Zhou M, Cai Y, Zhang Y, Pan X. 2011. Does temperature enhance acute mortality effects of ambient particle pollution in Tianjin City, China. Sci Total Environ 409:1811-1817.

Li N, Hao M, Phalen RF, Hinds WC, Nel AE. 2003. Particulate air pollutants and asthma. A paradigm for the role of oxidative stress in PM-induced adverse health effects. Clin Immunol 109:250-265.

Lim SS, Vos T, Flaxman AD, Danaei G, Shibuya K, Adair-Rohani H, et al. 2012. A comparative risk assessment of burden of disease and injury attributable to 67 risk factors and risk factor clusters in 21 regions, 1990-2010: a systematic analysis for the Global Burden of Disease Study 2010. Lancet 380:2224-2260.

Michelozzi P, Accetta G, De Sario M, D'Ippoliti D, Marino C, Baccini M, et al. 2009. High temperature and hospitalizations for cardiovascular and respiratory causes in 12 European cities. Am J Respir Crit Care Med 179:383-389.

Mirabelli MC, Beavers SF, Chatterjee AB. 2014. Active asthma and the prevalence of physician-diagnosed COPD. Lung 192:693-700.

Mirabelli MC, Vaidyanathan A, Qin X, Garbe PL. 2015. County-level PM_{2.5} and asthma symptoms in the past 14 days among U.S. adults with active asthma [Abstract]. Am J Respir Crit Care Med 191:A6264.

Mitchell KE, Lohmann D, Houser PR, Wood EF, Schaake JC, Robock A, et al. 2004. The multi-institution North American Land Data Assimilation System (NLDAS): Utilizing multiple GCIP products and partners in a continental distributed hydrological modeling system. J Geophysical Research 109:D07S90.

Natarajan S, Lipsitz SR, Fitzmaurice G, Moore CG, Gonin R. 2008. Variance estimation in complex survey sampling for generalized linear models. J Roy Stat Soc C-App 57:75-87.

National Asthma Control Program. 2011a. 2009 Behavioral Risk Factor Surveillance System Asthma Call-Back Survey Summary Data Quality Report. Available: http://www.cdc.gov/brfss/acbs/2009/documentation/SDQReportACBS_09.rtf [accessed August 26 2013].

National Asthma Control Program. 2011b. 2006-2008 Behavioral Risk Factor Surveillance System Asthma Call-Back Survey Summary Data Quality Report. Available: http://www.cdc.gov/brfss/acbs/2008/documentation/SDQReportACBS_06-08.rtf [accessed August 26 2013].

National Asthma Control Program. 2012. 2010 Behavioral Risk Factor Surveillance System Asthma Call-back Survey Summary Quality Data Report. Available: http://www.cdc.gov/brfss/acbs/2010/documentation/SDQReportACBS_10.rtf [accessed March 11 2013].

National Climatic Data Center. National Oceanic and Atmospheric Administration. U.S. climate regions. Available: http://www.ncdc.noaa.gov/monitoring-references/maps/us-climate-regions.php [accessed September 10 2014].

Nawrot TS, Torfs R, Fierens F, De Henauw S, Hoet PH, Van Kersschaever G, et al. 2007. Stronger associations between daily mortality and fine particulate air pollution in summer than in winter: evidence from a heavily polluted region in western Europe. J Epidemiol Community Health 61:146-149.

Nel AE, Diaz-Sanchez D, Li N. 2001. The role of particulate pollutants in pulmonary inflammation and asthma: evidence for the involvement of organic chemicals and oxidative stress. Curr Opin Pulm Med 7:20-26.

Park AK, Hong YC, Kim H. 2011. Effect of changes in season and temperature on mortality associated with air pollution in Seoul, Korea. J Epidemiol Community Health 65:368-375.

Pope CA, 3rd, Kalkstein LS. 1996. Synoptic weather modeling and estimates of the exposure-response relationship between daily mortality and particulate air pollution. Environ Health Perspect 104:414-420.

Ren C, Williams GM, Tong S. 2006. Does particulate matter modify the association between temperature and cardiorespiratory diseases? Environ Health Perspect 114:1690-1696.

Samet J, Zeger S, Kelsall J, Xu J, Kalkstein L. 1998. Does weather confound or modify the association of particulate air pollution with mortality? An analysis of the Philadelphia data, 1973-1980. Environ Res 77:9-19.

Samet JM, Dominici F, Curriero FC, Coursac I, Zeger SL. 2000. Fine particulate air pollution and mortality in 20 U.S. cities, 1987-1994. N Engl J Med 343:1742-1749.

Slade GD, Sanders AE, By K. 2012. Role of allostatic load in sociodemographic patterns of pain prevalence in the U.S. population. J Pain 13:666-675.

U.S. Department of Agriculture. 2013. Rural-urban continuum codes: documentation. Available: http://www.ers.usda.gov/data-products/rural-urban-continuum-codes/documentation.aspx [accessed September 10 2014].

U.S. Department of Commerce Bureau of the Census. 1994. Geographic Areas Reference Manual, Chapter 4: States, Counties, and Statistically Equivalent Entities. Available: http://www2.census.gov/geo/pdfs/reference/GARM/GARMcont.pdf [accessed February 4, 2016].

U.S. Environmental Protection Agency. Technology Transfer Network (TTN) Air Quality System (AQS). Available: http://www.epa.gov/ttn/airs/airsaqs/basic_info.htm [accessed September 11 2014].

Vahlsing C, Smith KR. 2012. Global review of national ambient air quality standards for PM₁₀ and SO₂ (24 h). Air Qual Atmos Health 5:393-399.

Vaidyanathan A, Dimmick WF, Kegler SR, Qualters JR. 2013. Statistical air quality predictions for public health surveillance: evaluation and generation of county level metrics of PM_{2.5} for the environmental public health tracking network. Int J Health Geogr 12:12.

Young MT, Sandler DP, DeRoo LA, Vedal S, Kaufman JD, London SJ. 2014. Ambient air pollution exposure and incident adult asthma in a nationwide cohort of U.S. women. Am J Respir Crit Care Med 190:914-921.

Zanobetti A, Schwartz J. 2009. The effect of fine and coarse particulate air pollution on mortality: a national analysis. Environ Health Perspect 117:898-903.

Table 1. Characteristics of the study sample and the population estimate, with percentages reporting asthma symptoms in the past 14 days

	Survey	Weig	thted population	Asthma
	sample		estimate	symptoms ^b
Characteristic	N	N ^a	Percent (95% CI)	Percent \pm SE
Total	50,356	17,963	,	57.1 ± 0.5
Age, in years	•	ŕ		
18–34	5,430	5,597	31.2 (30.0, 32.4)	51.1 ± 1.3
35–44	6,653	3,318	18.5 (17.7, 19.3)	57.1 ± 1.2
45–54	11,237	3,515	19.6 (18.9, 20.2)	61.0 ± 0.9
55–64	13,075	2,833	15.8 (15.2, 16.3)	61.2 ± 0.8
65–99	13,961	2,701	15.0 (14.5, 15.5)	60.4 ± 0.8
Educational attainment	•	ŕ		
Less than high school	4,795	1,817	10.1 (9.5, 10.7)	63.2 ± 1.6
Graduated high school	13,372	4,628	25.8 (24.8, 26.8)	60.2 ± 1.2
College 1–3 years/technical school	15,062	5,120	28.5 (27.6, 29.4)	59.5 ± 1.0
College 4+ years	17,127	6,398	35.6 (34.6, 36.6)	51.4 ± 0.8
Race/ethnicity				
White, non-Hispanic	42,143	13,415	74.7 (73.6, 75.7)	58.3 ± 0.6
Black, non-Hispanic	3,105	1,656	9.2 (8.6, 9.9)	55.4 ± 1.9
Other, non-Hispanic	2,812	1,161	6.5 (5.9, 7.1)	61.8 ± 2.5
Hispanic	2,296	1,731	9.6 (8.8, 10.5)	46.8 ± 2.4
Sex				
Female	36,995	11,244	62.6 (61.5, 63.7)	58.0 ± 0.6)
Male	13,361	6,719	37.4 (36.3, 38.5)	55.8 ± 1.0)
Smoking status				
Current smoker	9,261	3,473	19.3 (18.5, 20.1)	69.8 ± 1.1)
Former smoker	17,085	4,877	27.1 (26.3, 28.0)	59.6 ± 0.9)
Lifetime non-smoker	24,010	9,613	53.5 (52.5, 54.5)	51.3 ± 0.8)
U.S. climate region				
Central	5,901	2,687	15.0 (14.4, 15.5)	60.8 ± 1.2)
East north central	4,575	1,401	7.8 (7.5, 8.1)	61.2 ± 1.2)
Northeast	13,481	4,582	25.5 (24.8, 26.3)	56.2 ± 1.1)
Northwest	5,943	741	4.1 (4.0, 4.3)	59.6 ± 1.2)
South	6,478	2,226	12.4 (11.8, 13.0)	62.1 ± 1.6)
Southeast	3,442	2,423	13.5 (12.9, 14.1)	59.0 ± 1.6)
Southwest	3,778	1,029	5.7 (5.4, 6.0)	60.7 ± 1.8)
West	2,587	2,676	14.9 (14.2, 15.6)	45.0 ± 1.7)
West north central	4,171	198	1.1 (1.0, 1.2)	57.9 ± 1.5
Urbanicity				
Rural	7,329	1,007	5.6 (5.3, 5.9)	63.1 ± 1.5)
Suburban	9,238	2,122	11.8 (11.3, 12.4)	65.1 ± 1.2)
Urban	33,789	14,834	82.6 (82.0, 83.2)	55.6 ± 0.6)

^a In thousands ^b In the past 14 days

Table 2. Summary statistics and correlation coefficients describing the distributions of PM_{2.5}, O₃, precipitation, and temperature

		Average	Percentiles ^a		Pearson's correlation coefficients ^b				
Exposure	N	daily mean \pm SD ^a	25^{th}	50 th	75 th	PM _{2.5}	O_3	Precipitation	Temperature
$PM_{2.5}(\mu g \cdot m^{-3})$	50,356	9.4 ± 3.1	7.1	9.0	11.4	1	0.03	-0.02	-0.02
O_3 (ppb)	50,356	38.7 ± 9.7	31.2	38.6	45.7		1	-0.11	0.57
Precipitation (mm)	50,356	2.8 ± 3.3	0.8	2.0	3.7			1	-0.04
Temperature (°F)	50,356	62.9 ± 19.3	48.5	64.6	77.9				1

SD, standard deviation; $PM_{2.5}$, particulate matter \leq 2.5 microns in diameter; O_3 , ozone ^a Based on unweighted survey data

Table 3. Associations between PM_{2.5} and the prevalence of asthma symptoms in the past 14 days among adults with active asthma

	Survey sample	Weighted population estimate	Unadjusted model	Partially mod	adjusted dels	Fully adjusted, final model	
Metrics of PM _{2.5}	N	N^a	PD (95% CI)	PD (95% CI) ^b	PD (95% CI) ^c	PD (95% CI) ^{b,c}	
Quartiles							
4.00–7.06 μg·m ⁻³	12,640	2,747	0.0 (referent)	0.0 (referent)	0.0 (referent)	0.0 (referent)	
$7.07-8.97 \mu \text{g} \cdot \text{m}^{-3}$	12,586	4,054	4.2 (1.1, 7.3)	4.2 (1.2, 7.1)	4.5 (1.4, 7.6)	4.4 (1.4, 7.4)	
$8.98-11.36 \mu \text{g} \cdot \text{m}^{-3}$	12,595	5,160	4.1 (1.0, 7.1)	4.3 (1.4, 7.2)	4.6 (1.4, 7.8)	4.7 (1.6, 7.8)	
11.37–19.98 μg·m ⁻³	12,535	6,002	2.5 (-0.6, 5.6)	3.0 (0.1, 5.9)	4.8 (1.3, 8.3)	4.9 (1.5, 8.2)	
Linear spline segments							
per $\mu g \cdot m^{-3} 4.00 - 7.06$	12,640	2,747	2.9 (0.6, 5.2)	3.1 (0.8, 5.3)	3.4 (1.1, 5.7)	3.4 (1.1, 5.7)	
per $\mu g \cdot m^{-3} 7.07 - 8.97$	12,586	4,054	0.9 (-1.4, 3.2)	0.8 (-1.3, 3.0)	0.9 (-1.4, 3.1)	0.9(-1.3, 3.0)	
per $\mu g \cdot m^{-3} 8.98 - 11.36$	12,595	5,160	-0.6 (-2.2, 1.1)	-0.5 (-2.0, 1.1)	0.0 (-1.7, 1.6)	-0.1 (-1.6, 1.5)	
per μg·m ⁻³ 11.37–19.98	12,535	6,002	-0.2 (-1.1, 0.8)	-0.1 (-1.0, 0.8)	0.2 (-0.7, 1.2)	0.3 (-0.6, 1.1)	
Continuous measure							
per $\mu g \cdot m^{-3} 4.00 - 19.98$	50,356	17,963	0.2 (-0.2, 0.6)	0.2 (-0.1, 0.5)	0.5 (0.1, 0.9)	0.5 (0.1, 0.9)	

CI, confidence interval; O₃, ozone; PD, percent difference

^a In thousands

^b Adjusted for individual-level covariates: age, educational attainment, race, sex, and smoking status

^c Adjusted for county-level covariates: O₃, precipitation, region, temperature, and urbanicity

FIGURE CAPTIONS

Figure 1. Weighted estimates of the population with active asthma (○) and estimated percentages of adults with asthma symptoms in the past 14 days (→) across categories of PM_{2.5} (panel A), ozone (panel B), precipitation (panel C), and temperature (panel D).

Figure 2. Associations between PM_{2.5} and the prevalence of asthma symptoms in the past 14 days among adults with active asthma across quintiles of air temperature. Results are shown for temperature category-stratified models in which PM_{2.5} is parameterized as quartiles (panel A), four linear spline segments (panel B), and a single continuous measure (panel C). All models are adjusted for individual- and county-level covariates.

Figure 1.

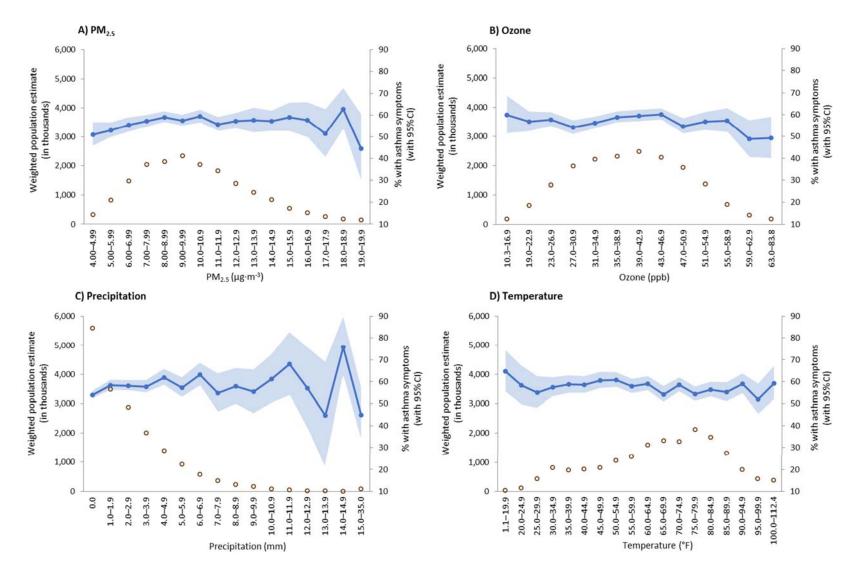
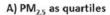
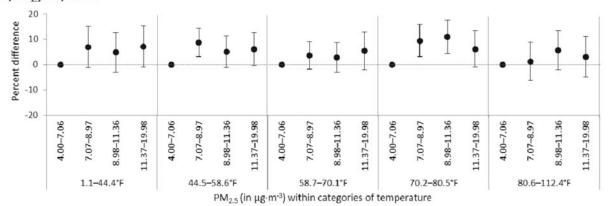
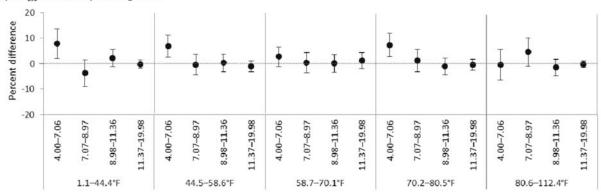


Figure 2.





B) PM_{2.5} as linear spline segments



PM_{2.5} (in μg·m⁻³) within categories of temperature

C) PM_{2.5} as a single continuous measure

